Changes in Fine PM Pollution Levels with Tightening of Regulations on Vehicle Emissions

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Abstract—A long-term campaign for monitoring the concentration of atmospheric Particulate Matter (PM) was conducted at multiple sites located in the center and suburbs of the Tokyo Metropolitan Area in Japan. The concentration of fine PM has shown a declining trend over the last two decades. A positive matrix factorization model elucidated that the contribution of combustion sources was drastically reduced. In Japan, the regulations on vehicle exhaust emissions were phased in and gradually tightened over the last two decades, which has triggered a notable reduction in PM emissions from automobiles and has contributed to the mitigation of the problem of fine PM pollution.

Keywords—Air pollution, Diesel-powered vehicle, Positive matrix factorization, Receptor modeling.

I. INTRODUCTION

RECENTLY, particulate matter (PM) has become one of the air pollution issues of greatest concern. In Japan, guidelines for the concentration of suspended particulate matter (SPM, representing PM less than or equal to 10 μm in aerodynamic diameter under the 100% cut-off condition) have been enforced since 1973. Owing to the tightening of regulations on several emission sources, the current SPM concentration has decreased notably [1]. Nonetheless, in the past several decades fine PM has been conclusively demonstrated to strongly affect human health [2]. In Japan, new guidelines for PM$_{2.5}$ ($35 \mu g/m^3$ 24-h mean and $15 \mu g/m^3$ annual mean) were introduced in 2009 [3].

In 1989, local governments in the center and suburbs of the Tokyo Metropolitan Area started a joint research project that aims to clarify the major sources of fine PM (PM$_{2.1}$) pollution [4], with particular focus on the Tokyo Metropolitan Area, which is characterized by a large population, heavy traffic, and massive industrial complexes. In the last several decades, regulations on vehicle emissions have been phased in and gradually tightened, and regulations for the transport sector tend to become tighter as a global trend.

Prior to investigating the introduction of further regulations, it is important to evaluate the effects of previous policies. By using positive matrix factorization (PMF), this study discusses whether previously enforced regulations on vehicle emissions have contributed to alleviating the fine PM pollution problem in the Tokyo Metropolitan Area.

II. EXPERIMENTAL

A. Sample Collection and Chemical Analysis

The Tokyo Metropolitan Area, covering a region with a radius of 70 km from the center of Tokyo, is characterized by a large population (more than 34 million people, corresponding to 30% of the total population of Japan) and heavy traffic (18 million vehicles corresponding to 20% of the total number of vehicles in Japan). Massive industrial complexes, including facilities operated by the steel and petrochemical industries, are located in Tokyo Bay.

A PM monitoring campaign was conducted between 1989 and 2007. Sample collection was carried out simultaneously at multiple sites (17 sites in 2007 (Fig. 1)) every summer and winter. PM samples categorized according to size (PM$_{2.1}$ and PM$_{2.1-11}$) were collected for 1 week by using an Andersen-type low-volume air sampler.

Carbon components (organic carbon (OC) and total carbon (TC)) were analyzed by using a thermal carbon analyzer. OC was separated under at 600 °C in a helium atmosphere, and total carbon (TC) was separated at 950 °C in a 2% oxygen/98% helium atmosphere. Elemental carbon (EC) was defined as the amount of OC subtracted from the amount of TC.

Ion species (Na$^+$, Mg$^{2+}$, K$^+$, Ca$^{2+}$, NH$_4^+$, Cl$^-$, NO$_3^-$, and SO$_4^{2-}$) were analyzed by using ion chromatography. A portion of a sample-loaded filter was immersed into 20 mL of ultrapure water and sonicated for 15 min in an ultrasonic bath, and the resulting solution was filtered through a PTFE membrane filter.

The presence of various elements (V, Al, Mn, Cl, Na, Ca, Br, Ti, Mg, and Cu) was determined by instrumental neutron
activation analysis. A portion of a sample-loaded filter was placed in a polyethylene bag and irradiated with thermal neutrons for 40 s, after which the gamma-ray spectra of short-lived radioisotopes were shortlisted.

The presence of polycyclic aromatic hydrocarbons (PAHs: B[k]F, B[a]P, and B[ghi]P) was determined by using high-performance liquid chromatography. A portion of a sample-loaded filter was immersed into 5 mL of dichloromethane and sonicated for 15 min in an ultrasonic bath. The resulting solution was filtered by using a PTFE membrane filter, and the filtrate was subsequently concentrated by blowing N2 and dissolved in 1.5 mL of acetonitrile.

B. Principle of PMF Model

The PMF model can extract several factors from observational datasets according to similarities in their temporal or spatial characteristics [5]. The model essentially assumes that the number of factors is p, in other words, several sources impacting a receptor, and that the linear combination of the impacts from these sources results in the observed concentrations of PM components [6]. The model can be expressed as in (1).

\[ x_{ij} = \sum_{k=1}^{p} g_{ik} f_{kj} + e_{ij} \]  

where \( i = 1, \ldots, n \) is the number of samples; \( j = 1, \ldots, m \) is the number of species; and \( k = 1, \ldots, p \) is the number of factors. Furthermore, \( x_{ij} \) is the observed concentration for the \( j \)th species, \( f_{kj} \) is the chemical profile of the \( k \)th factor for the \( j \)th species, and \( e_{ij} \) is the residual for the \( j \)th species in the \( i \)th sample. In PMF, the objective is to estimate the contributions of factors to the observed concentrations in the \( i \)th sample. The requirement of PMF is to minimize the objective function \( Q \), which is expressed as in (2).

\[ Q = \sum_{i=1}^{n} \sum_{j=1}^{m} \left( \frac{e_{ij}^2}{u_{ij}} \right)^2 = \sum_{i=1}^{n} \sum_{j=1}^{m} \left( x_{ij} - \sum_{k=1}^{p} g_{ik} f_{kj} \right)^2 \]  

where \( u_{ij} \) is the uncertainty with respect to the \( j \)th species in the \( i \)th sample. The uncertainties for the respective observational data were calculated by using (3) in the case of \( x_{ij} \leq MDL_j \) and (4) in the case of \( x_{ij} > MDL_j \) [7].

\[ u_{ij} = \frac{5}{6} \times MDL_j \]  

\[ u_{ij} = \sqrt{(Error\ Frac.\times x_{ij})^2 + MDL_j^2} \]  

where \( MDL_j \) is the method detection limit in the \( j \)th species and \( Error\ Frac. \) is the percentage of uncertainty for the analytical value of the \( j \)th species. In this study, \( Error\ Frac. \) was set to 10%. Moreover, the uncertainties for each species were appropriately corrected according to the signal-to-noise (S/N) ratio, which is expressed as in (5) [8].

\[ \left( \frac{S}{N} \right)_j = \frac{\sum_{i=1}^{n} (x_{ij} - u_{ij})^2}{\sum_{i=1}^{n} u_{ij}^2} \]  

In order to minimize the impact on the PMF solution, the uncertainties for species with an S/N ratio of less than 2 were corrected to 3-fold of original values (set to "weak" in the model). The mathematical algorithm was processed with the EPA-PMF 3.0 software.

III. RESULTS AND DISCUSSION

A. Dataset for PMF Modeling

To create the dataset for PMF modeling, we selected 16 species (OC, EC, Na+, Mg2+, K+, Ca2+, NH4+, Cl–, NO3-, SO42-, Al, V, Mn, B[k]F, B[a]P, and B[ghi]P), whose concentrations were continuously monitored during the 19-year period of the monitoring campaign. The dataset (\( x_{ij} \) matrix) includes approximately 20,000 entries of observation data. Table 1 summarizes the statistics of the dataset for PMF modeling.

<table>
<thead>
<tr>
<th>Species</th>
<th>Unit</th>
<th>MDL</th>
<th>Concentration</th>
<th># Samples</th>
</tr>
</thead>
<tbody>
<tr>
<td>OC</td>
<td>µg/m³</td>
<td>0.10</td>
<td>Mean: 3.93, Max: 22.3, Min: 0.12</td>
<td>0</td>
</tr>
<tr>
<td>EC</td>
<td>µg/m³</td>
<td>0.05</td>
<td>Mean: 2.62, Max: 14.8, Min: 0.07</td>
<td>6</td>
</tr>
<tr>
<td>Na+</td>
<td>µg/m³</td>
<td>0.01</td>
<td>Mean: 0.44, Max: 7.2, Min: 0.02</td>
<td>5</td>
</tr>
<tr>
<td>Mg2+</td>
<td>µg/m³</td>
<td>0.005</td>
<td>Mean: 0.06, Max: 0.83, Min: 0.01</td>
<td>49</td>
</tr>
<tr>
<td>K+</td>
<td>µg/m³</td>
<td>0.005</td>
<td>Mean: 0.14, Max: 0.99, Min: 0.01</td>
<td>3</td>
</tr>
<tr>
<td>Ca2+</td>
<td>µg/m³</td>
<td>0.01</td>
<td>Mean: 0.33, Max: 2.87, Min: 0.02</td>
<td>22</td>
</tr>
<tr>
<td>NH4+</td>
<td>µg/m³</td>
<td>0.005</td>
<td>Mean: 1.23, Max: 9.11, Min: 0.01</td>
<td>26</td>
</tr>
<tr>
<td>Cl–</td>
<td>µg/m³</td>
<td>0.005</td>
<td>Mean: 0.96, Max: 13.6, Min: 0.01</td>
<td>9</td>
</tr>
<tr>
<td>NO3-</td>
<td>µg/m³</td>
<td>0.01</td>
<td>Mean: 1.37, Max: 10.6, Min: 0.02</td>
<td>3</td>
</tr>
<tr>
<td>SO42-</td>
<td>µg/m³</td>
<td>0.03</td>
<td>Mean: 2.54, Max: 25.8, Min: 0.06</td>
<td>3</td>
</tr>
<tr>
<td>Al</td>
<td>µg/m³</td>
<td>0.005</td>
<td>Mean: 0.28, Max: 2.84, Min: 0.01</td>
<td>12</td>
</tr>
<tr>
<td>V</td>
<td>mg/m³</td>
<td>0.5</td>
<td>Mean: 3.53, Max: 35.0, Min: 1.0</td>
<td>84</td>
</tr>
<tr>
<td>Mn</td>
<td>mg/m³</td>
<td>5.0</td>
<td>Mean: 18.3, Max: 160, Min: 10.0</td>
<td>269</td>
</tr>
<tr>
<td>B[k]F</td>
<td>ng/m³</td>
<td>0.005</td>
<td>Mean: 0.28, Max: 4.18, Min: 0.01</td>
<td>81</td>
</tr>
<tr>
<td>B[a]P</td>
<td>ng/m³</td>
<td>0.005</td>
<td>Mean: 0.34, Max: 5.37, Min: 0.01</td>
<td>64</td>
</tr>
<tr>
<td>B[ghi]P</td>
<td>ng/m³</td>
<td>0.005</td>
<td>Mean: 0.65, Max: 13.2, Min: 0.01</td>
<td>106</td>
</tr>
</tbody>
</table>

B. Temporal Trends for PM2.1 and EC

Fig. 2 shows the temporal trends in averaged mass concentrations of PM2.1 and EC for all sites in the winter campaign. The concentrations of PM2.1 and EC show clear temporal trends can be seen in the cases of OC, NO3–, Cl–, and PAHs. On the other hand, there are no notable changes in the concentrations of the other species. This study focuses on the reasons for this remarkable reduction in PM2.1 concentration in winter.
C. Identification of PM$_{2.5}$ Sources and Trends in the Contribution from Automotive Exhaust Products

Preliminary cluster analysis has suggested that between 5 and 7 factors might be extractable from the dataset. We extracted 6 base factors on the basis of the following criteria:

- Ensuring conformance of Q to the theoretical value.
- Maintaining the ion balance in the extracted factors.
- Reproducing the concentrations of 16 species.

Significant positive correlation ($r = 0.98$, $p < 0.01$) was found between the observations and calculated values for 6 base factors. A bootstrap run confirmed the stability of the base run solutions. Moreover, no significant rotation freedom was found in the base run solutions. Consequently, we adopted these 6 base factors as final solutions.

Fig. 3 shows the factor profiles as a distribution of chemical species across these 6 factors. By focusing on the indicative species, the potential sources of the respective factors were interpreted as follows.

- Factor 1 includes NH$_4^+$ and SO$_4^{2-}$, representing the impact of secondary sulfates (ammonium sulfate). In addition, OC, EC, K$^+$, and V, which are mainly emitted from combustion sources, can also be associated with this factor. Consequently, we interpret Factor 1 as secondary sulfates, including the impact of combustion sources as the origin of precursors of secondary sulfates.
- Factor 2 includes OC, EC, and PAHs. These are species indicative of combustion sources.
- Factor 3 includes NH$_4^+$, Cl$^-$, and NO$_3^-$, representing the impacts of secondary chlorides (ammonium chloride) and secondary nitrates (ammonium nitrate).
- Factor 4 includes Ca$^{2+}$ and Al, which are emitted mainly as a result of weathering of geological media. In this study, we interpret Factor 4 as road dust.
- Factor 5 includes Na$^+$, Mg$^{2+}$, and NO$_3^-$ as indicative species. Sodium nitrate or magnesium nitrate can be formed by a heterogeneous reaction between sea salt particles and HNO$_3$ gas. Therefore, Factor 5 represents aged sea salt.
- Factor 6 includes Na$^+$, Mg$^{2+}$, and Cl$^-$, representing the impact of fresh sea salt.

Factors 1, 2, and 3 contribute strongly to the PM$_{2.5}$ fraction, suggesting that these factors originate mainly from anthropogenic sources, whereas Factors 4, 5, and 6 are typically associated with the PM$_{2.5-11}$ fraction, indicating that they originate from natural sources. These features are consistent with the above interpretations.

Fig. 4 shows the variation in averaged factor contribution to the PM$_{2.5}$ concentration in the winter campaign. As mentioned above, the averaged mass concentration of PM$_{2.5}$ has shown a trend toward decrease over the last two decades, and Factor 2 contributes notably to this trend. As shown in Fig. 3, Factor 2 includes OC, EC, and PAHs as indicative species, which are mainly emitted from combustion sources such as internal combustion engines, biomass burning, and oil burning. There is a fair positive correlation ($r = 0.90$, $p < 0.01$) between the EC concentration as calculated from Factor 2 and the observed EC concentration. Therefore, the reduction in EC concentration can be attributed to the reduction of PM emissions from combustion sources.
Focusing on the recently implemented regulations on vehicle emissions, as shown in Fig. 5, regulations on vehicle exhaust emissions have been phased in and gradually tightened over the last two decades. With respect to the regulations on new cars, the Japanese government introduced new regulations on emission standards for NOx and PM in 1994, 1998, 2003, and 2005 (corresponding to (A), (B), (D), and (E) in Fig. 5, respectively). These regulations resulted in the reduction of the emission of air pollutants from vehicles, particularly from diesel-powered vehicles. This drastic change was most likely due to the Law Concerning Special Measures for Total Emission Reduction of Nitrogen Oxides and Particulate Matter (corresponding to (C) in Fig. 5), which was enforced in July 2001. In accordance with national legislation aimed at curbing NOx and PM emissions in major metropolitan areas, inspection compliance cannot be issued for vehicles that fail to meet the legal standards at inspection time. The enforcement of these regulations means that restrictions are imposed on diesel-powered vehicle circulation in the areas concerned.

The following regulations were introduced: (A) Short-term automobile exhaust gas regulations (1994), (B) Long-term automobile exhaust gas regulations (1998), (C) Law Concerning Special Measure for Total Emission Reduction of Nitrogen Oxides and Particulate Matter (2001), (D) New short-term automobile exhaust gas regulations (2003), and (E) New long-term automobile exhaust gas regulations (2005). The monitoring campaign was divided into 6 intervals ((P1)–(P6)) associated with the tightening of the regulations.

To estimate the effects of these regulations, we divided the 19-year period of the monitoring campaign into 6 intervals associated with the tightening of regulations (corresponding to (P1)–(P6) in Fig. 5), and Fig. 6 shows the contribution of Factor 2 to PM$_{2.5}$ concentration in the respective intervals. During the first decade of the campaign, the contribution of Factor 2 to PM$_{2.5}$ concentration was greater than 30%; however, phased tightening of the regulations on vehicle emissions has clearly triggered a notable reduction in the contribution from Factor 2, which is currently approximately 10%. In conclusion, the tightening of the regulations on vehicle emissions has decidedly contributed to the mitigation of the fine PM pollution in the center and suburbs of the Tokyo Metropolitan Area.

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### REFERENCES